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**Entrainment and Optical Properties of an Elevated Canadian Forest Fire Plume
Transported into the Planetary Boundary Layer near Washington, D.C.**

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Transported into the Planetary Boundary Layer near Washington, D.C.**

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Abstract

Smoke and pollutants from Canadian forest fires were transported over the northeastern United States in July 2002. Lidar observations at the NASA Goddard Space Flight Center show the smoke from these fires arriving in an elevated plume that was subsequently mixed to the surface. Trajectory and three-dimensional model calculations confirm the origin of the smoke and show that it mixed to the surface after it was intercepted by the turbulent planetary boundary layer. Modeled smoke optical properties agreed well with aircraft and remote sensing observations provided coagulation of smoke particles was accounted for in the model. Our results have important implications for the long-range transport of pollutants and their subsequent entrainment to the surface, as well as the evolving optical properties of smoke from boreal forest fires.

Introduction

Ozone and aerosols transported over long distances can affect air quality at local and regional scales. For example, the air over the Mediterranean contains a complicated mix of pollutants transported from Asia, North America, and Europe (1). Saharan dust transported to the U.S. in the summertime occasionally contributes enough dust to put portions of Florida out of compliance with U.S. Environmental Protection Agency (EPA) standards for fine particulate matter (PM_{2.5}, or particulate matter less than 2.5 μm diameter) (2). High surface-level concentrations of carbon monoxide (CO) observed during summer 1995 in the eastern and southeastern U.S. were attributed to pollutants produced and transported in the plumes from large Canadian forest fires (3). CO and nitrogen oxide (NO, also produced in the forest fires) are important precursors to the photochemical production of tropospheric ozone. Pollutants from Canadian and other northern forest fires are frequently transported long distances at low altitudes behind advancing cold fronts. In this paper, however, we discuss a case in which smoke from a Canadian forest fire was observed at the NASA Goddard Space Flight Center (GSFC, 39.02° N, 76.86° W, near Washington, D.C.) in an elevated plume that was subsequently mixed to the surface when it was intercepted by the turbulent planetary boundary layer (PBL).

Over the period July 5 – 9, 2002, lightning initiated multiple fires in central Quebec which burned about 250,000 ha (1 ha = 10^4 m^2) of boreal forest (4). An upper level low-pressure system over the Canadian Maritime Provinces coupled with a high pressure ridge to the west channeled the smoke and pollutants from the fires about 1000 km

southward to the eastern U.S. These plumes were clearly visible in satellite imagery (Figure S1), and aerosol optical thickness (AOT) retrievals from the Earth-Probe Total Ozone Mapping Spectrometer (EP-TOMS) satellite instrument (5) showed that the plumes were optically thick (Figure 1). During the passage of the plumes the U.S. East Coast experienced some of the worst air quality days of the summer.

Evolution of Plume Vertical Structure

The NASA Micro-Pulse Lidar NETwork (MPLNET) operates a lidar at GSFC, co-located with a sun/sky photometer in the Aerosol Robotic NETwork (AERONET) (6,7). Figure 2a shows the MPLNET normalized relative backscatter ratio (8) measured by the lidar from July 6 – 9, 2002. These measurements show a smoke layer between 2 – 3 km arriving on July 6. On July 7 the smoke plume rapidly mixed to the surface after about 1200 UTC. Because the MPLNET signal is completely attenuated above the lowest part of the smoke plume we do not have any information on the thickness of the smoke layer on July 7. On July 8 there is still an elevated plume around 2 km. Also shown in Figure 2 is the PBL height at GSFC from the NCAR Model for Atmospheric Transport and Chemistry (MATCH) (9). The peak height of the MATCH PBL is as high as the top of the MPLNET signal at 1800 UTC on July 7, and the subsequent collapse of the model's PBL later in the day is coincident with the observed plume descent and mixing to the surface.

Trajectory analyses carried out with the NASA Goddard trajectory model (10) using the NCEP/NCAR reanalyses (11) show that the July 6 – 8 plume observed by MPLNET

over GSFC originated 24 – 48 hours earlier over the Quebec fire regions (Figure S2). Figure 3 shows two-day back-trajectories initialized at GSFC at the altitude of the MPLNET observed peak backscatter in the smoke plume. Trajectories were run for every hour between 1800 UTC July 6 (when the smoke arrives) and 1200 UTC July 8 (when the MPLNET lidar was shut off). In determining the altitude of the plume we only consider backscatter altitudes above 0.75 km so that we are not biased by low altitude anthropogenic pollutants. The times the fires were actively burning in Quebec were determined from GOES-8 geostationary satellite imagery (12) and are indicated by the grey bars in Figure 3. For each trajectory we show the time and the altitude it crosses southward of 51° N, roughly the southern edge of the fire region identified from the GOES-8 imagery. For almost all of the trajectories this point occurred during or shortly after the period when the fires were active on July 5. The trajectories were between 2 – 6 km altitude over the fire region and descended during subsequent transport. Trajectories at higher altitudes over the fire region were transported to GSFC more rapidly than those that were at lower altitudes.

On July 6 and 7 the surface ozone levels (13) were low over most of the northeastern U.S. (Figure S3). The pollutants from the fires had not yet been transported to the surface. Figure 4a shows that on July 8, however, there were very high surface mixing ratios of ozone (mixing ratio > 125 ppbv) near Washington, D.C., across New Jersey, and in Ohio and western Pennsylvania. The high surface ozone near GSFC on July 8 occurred shortly after MPLNET observations showed the pollutants from the fire plume mixing to the surface.

Modeled Smoke Plume

Based on these data and our trajectory calculations we hypothesize that the pollutants from the fires were transported from Canada toward GSFC in an elevated but subsiding plume, and that this plume was rapidly mixed to the surface over GSFC on July 7 as it was intercepted by the turbulent PBL. To test this hypothesis we simulated the evolution of the smoke plume with a three-dimensional Eulerian aerosol microphysical and transport model (14). The model dynamics are from the NCEP/NCAR reanalyses, with parameterizations for moist convection, precipitation, and boundary layer mixing from MATCH (15,16). Fire locations and timing were prescribed in the model using fire hot spots and emission times visually identified from GOES-8 imagery. The satellite imagery showed the fires had a distinct diurnal cycle (Figure 3). In these simulations we treat the smoke aerosol particle size distribution in 16 discrete size bins space between $0.01 - 1 \mu\text{m}$ radius. Because the aerosol particles are all sub-micron in size they are not strongly affected by sedimentation. Particles are removed from the model by precipitation scavenging (which can occur wherever there is precipitation) or by dry deposition (which occurs only at the surface), we determined that neither of these processes were especially important to our simulated smoke distributions over GSFC.

The modeled vertical distribution of smoke at GSFC is sensitive to the height at which we assume emissions from the fires occur. Global scale models typically inject biomass burning emissions near the surface or uniformly mix them throughout the PBL (17), but there is ample evidence that intense fires or fires occurring near convective cells

can loft smoke and pollutants to much higher altitudes (18). The trajectory model, as well as the MPLNET data and aircraft observations discussed later, suggest that the material arriving at GSFC was initially injected between 2 – 6 km, so in our baseline model run we add smoke to the model during the active fire times by distributing it uniformly in this altitude range over the fire hot spots. Figure 2b shows the resulting vertical smoke mass distribution from the model over GSFC. The model vertical profile is similar to the MPLNET backscatter observations (Figure 2a), showing the arrival of an elevated plume late on July 6, a subsequent increase in surface mass concentrations on July 7, and an elevated plume again on July 8. Because the MPLNET observations are attenuated above 2 km on July 7 we cannot use them to verify the existence of the higher altitude model plume that appears at 4 km late on July 7. Sensitivity tests show that this relatively isolated high altitude plume is from the fires burning on July 6, whereas the rest of the profile—including the elevated plume on July 8—can be explained almost entirely by the fires burning on July 5 (consistent with our trajectory analysis). The smoke plumes were continuously and clearly visible in GOES-8 daytime imagery. This imagery suggested that most of the smoke over GSFC during this period is from the fires on July 5, so it is unlikely the high altitude plume in the model on July 7 is real.

In order to understand the observed vertical profile at GSFC we considered several sensitivity tests of our model. In one case we injected all of the smoke at the surface. Because of moist convection over the fire regions on July 5, a large fraction of this smoke was lifted to 2 – 8 km. In this simulation the vertical profile over GSFC was similar to the baseline case because both simulations had a great deal of smoke in the 2 – 6 km

range. Because the moist convection was not as strong on July 6 the smoke injected remains near the surface and the model did not produce a high altitude plume over GSFC on July 7. The convective mixing parameterization in the model does not greatly redistribute material already above the surface and below the convective cloud top, so in our baseline simulation the plume over the fire region remained relatively intact. In another sensitivity test we considered injecting the material in a layer between 0.5 – 3 km altitude. In this case the aerosol descended to the surface rapidly due to subsidence and the vertical profile at GSFC did not agree well with the MPLNET observations. In summary, the modeled vertical profile was similar to the MPLNET observations only when some of the smoke was lifted to 2 – 6 km over the fire regions on July 5. Moist convection over the fires provides a plausible mechanism for elevating the smoke.

The above discussion largely explains the arrival of an elevated plume at GSFC. Because of subsidence the plume descends during transport, as shown in Figure 3, but subsidence is not sufficient to explain how the material gets to the surface at GSFC. We test the importance of PBL entrainment in the model by turning off the PBL mixing mechanism downwind of the fire regions. For this model run the vertical profile over GSFC looks similar to our baseline simulation shown in Figure 2b except that almost all of the aerosol remains above about 500 m (Figure S4). For the high altitude smoke injection case with PBL entrainment, the peak surface mass concentration over GSFC is about $140 \mu\text{g m}^{-3}$. When PBL entrainment is turned off the peak in the surface mass concentration at GSFC falls to about $4 \mu\text{g m}^{-3}$. Figure 4 shows the modeled surface mass concentration of the smoke aerosol over the northeastern U.S. for the baseline simulations

both with and without entrainment into the PBL. For the case with PBL entrainment the distribution of high surface mass loadings of aerosol is similar to the distribution of high ozone occurrences in the EPA data over Ohio, Pennsylvania, and New Jersey on July 8. With PBL entrainment turned off, however, this correlation is not apparent. We conclude then that aerosol from the fires was transported mainly in an elevated plume, and that entrainment into the PBL was critical to bringing the aerosol to the surface near GSFC. Other pollutants transported with the aerosol in the plume (e.g., CO, NO, ozone) would have been transported in a similar fashion.

July 8 Aircraft Profiles of Plume

The remnant smoke plume observed on July 8 by MPLNET at GSFC was profiled several times with the University Research Foundation-Advanced Development Laboratory's twin engine Piper Aztec-F PA-23-250 research aircraft. The Aztec was outfitted with an atmospheric research package run by the University of Maryland (19). Particle scattering measurements were made with a three-channel integrating nephelometer operating at 450, 550, and 700 nm. Aerosol absorption was measured with a particle/soot absorption photometer operating at 565 nm. Additional measurements made include particle number concentrations, and CO, O₃, and SO₂ mixing ratios. Five vertical profiles were made from near the surface to the aircraft's operational ceiling at 3 km. Four of these profiles were dominantly influenced by plume material from the Canadian fires. A fifth profile was flown to the west of the main smoke plume and was more strongly influenced by anthropogenic pollutants in the PBL. The profiles spanned a

region about 200 km east-west and occurred over a period of about 8 hours. There is some variability in the smoke aerosol properties owing to the time and space variations in the plume. For the four profiles which showed evidence of the smoke plume, the layer was always at about 2 – 3 km and had distinctly different scattering and absorption properties from the aerosols in the PBL. Back trajectories calculated from this 2 – 3 km range show that for each profile the air passed over the fire regions during the burning on July 5 at about 4 km. Finally, for all of these profiles the enhanced aerosol scattering and absorption were well correlated with enhanced particle number concentrations and CO and O₃ mixing ratios.

Figure 5 shows the narrow layer of aerosol scattering and absorption observed in the smoke plume on the Aztec profile flown above Easton, MD (38.80° N, 76.06° W, about 70 km E of GSFC, profile flown between 19:45 – 20:12 UTC). Also shown are the scattering and absorption profiles from the MPLNET data at GSFC at 11:35 UTC July 8 (determined from the retrieved extinction profile and the AERONET single scatter albedo in Table 1), which show a peak in the 2 – 3 km altitude range. The MPLNET data also show a strong scattering and absorption peak below 1 km. Because of the high extinction-to-backscatter ratio determined for the MPLNET lidar profile it is likely that this low altitude peak is due to remnants from the smoke plume intercepted at the surface on July 7. Figure 5 also shows the modeled scattering and absorption profiles at GSFC at 2000 UTC on July 8. The model clearly shows an elevated smoke layer, although it is at a lower altitude (1.5 – 2.5 km) than the Aztec or MPLNET profiles. The peak magnitude of the modeled scattering and absorption are smaller than the lidar or aircraft

measurements, but the vertical extent of the smoke is much greater in the model than in the observations. The model vertical grid spacing varies from 100 – 200 m below 1 km altitude, and increases to about 500 m at 2 km and 800 m at 5 km. The vertical grid spacing is thus too coarse to adequately resolve the sharp peaks seen with the aircraft and lidar.

Smoke Plume Optical Properties

The data from the AZTEC profiles, along with the AERONET measurements at GSFC and the observations from the EP-TOMS space-based sensor, provide important constraints on the downwind optical properties of smoke aerosols from boreal forest fires. Integrating all of these observations into our aerosol model would be a formidable challenge, requiring a highly detailed characterization of the timing of the fires, the various materials burned, and the composition of the aerosol particles produced. There are additionally transformational processes that would affect the aerosol particle size distribution during transport. Previous studies on the aging of smoke aerosols have determined that coagulation is important in modifying the smoke particle size distribution (20,21). This process is generally neglected in global modeling studies because of the computational expense of carrying out the calculation. We have run simulations both with and without coagulation in order to investigate its importance in modifying the aerosol particle size distribution. For both simulations we initialize the model with a particle size distribution representative of fresh boreal smoke emissions (see supplemental online material). Without coagulation, the particle size distribution is

modified only slightly by removal processes. With coagulation the particle size distribution shifts to larger particles as expected.

Table 1 summarizes the observed and modeled optical properties of the smoke plume near GSFC on July 8. The EP-TOMS retrieved AOT is taken at 16:11 UTC, which is the closest satellite footprint to GSFC, and is slightly larger than the AERONET 380 nm measurements occurring within 30 minutes of the satellite overpass. The AERONET measurements at longer wavelengths are comparable to the column AOT measurements from the Aztec profiles of the smoke plume. The model overestimates the AOT slightly relative to the AERONET and Aztec measurements, but is comparable to EP-TOMS.

The single scatter albedo from the AERONET and Aztec measurements indicate a weakly absorbing aerosol. The model single scatter albedo agrees best with the AERONET measurements. The lower single scatter albedo determined from EP-TOMS is because the optical effective radius of the smoke particle size distribution assumed in the retrieval is too small compared to the AERONET measurements of this plume. Underestimating the particle effective radius makes the aerosol look more absorbing than it actually is, but has only a small effect on the retrieved AOT (5).

Scattering Ångström exponents determined inside the smoke plume from the Aztec nephelometer measurements (19) are recorded for each wavelength pair in Table 1. These parameters give information about the aerosol particle size distribution. For the model run with particle coagulation the Ångström exponents calculated at GSFC on July 8 are similar to the values determined from the aircraft measurements. Without coagulation, however, the model's Ångström exponents are much larger than and outside

the error bars on the observations. Additionally, the modeled particle size distribution at GSFC agrees well with AERONET retrievals of the particle size distribution when coagulation is accounted for and substantially underestimates the mean particle radius when coagulation is ignored (Figure S5).

There are uncertainties in comparing the model to these measurements. Because of its coarse spatial resolution it is difficult to justify comparing the model at the scale of a single grid-column to the essentially point-like measurements from the aircraft profiles and the sunphotometer measurements. On the other hand, the spatial extent of the plume is comparable to the model grid size. The model vertical profiles are similar at surrounding grid points, although there is some variability in both the magnitude and spectral dependence of the AOT. Since the magnitude of the AOT is a measure of how much aerosol is present it is not surprising that the AOT spectral dependence should be well correlated with the magnitude of the AOT. Where the AOT is small there are fewer particles and coagulation has been less important in modifying the particle size distribution, leading to larger Ångström exponents.

Despite these uncertainties our results lead us to conclude that coagulation is important to modifying the initial aerosol particle size distribution and can explain at least some of the observed spectral dependence in the AOT far downwind. This has implications to radiative forcings calculated with similar atmospheric transport models as well as retrievals of aerosol properties made from satellite observations. Neglecting the growth of aerosol particles in smoke plumes (“aging”) and instead relying on particle size distributions measured near the smoke emission point for these calculations will yield the

wrong spectral dependence in the optical properties.

Conclusions

Our analysis of the transport of a Canadian forest fire plume over the eastern U.S. during July 2002 demonstrates that an important mechanism for transporting pollutants from elevated layers to the surface is direct entrainment into the PBL of a gradually subsiding plume. An aerosol microphysics and transport model driven by assimilated meteorology confirms this mechanism. This model also tested various assumptions about the initial injection altitude of the plume from the fires. The modeled vertical profile at GSFC was similar to the MPLNET observations provided the majority of the smoke was initially injected into a 2 – 6 km altitude layer, as suggested by our trajectory calculations. Moist convection over the fire regions can explain how this elevated layer developed. Our simulation yielded spectral AOT values similar to those observed with the aircraft flights and sunphotometer measurements provided we accounted for modification of the smoke particle size distribution by coagulation. The aerosol lifting and deposition mechanisms discussed here have important implications for the long-range transport of smoke aerosols from boreal forest fires and their effects at the surface through subsequent coupling to the PBL downstream. Accounting for the modification of the smoker aerosol particle size distribution due to coagulation and other modification processes is also important for correctly determining the direct radiative effect of the aerosol, as well as for quantitative retrievals of aerosol properties from remote sensing measurements (22,23).

Figure 1. EP-TOMS aerosol optical thickness (380 nm) of smoke plume over northeastern U.S. on July 7, 2002. The dashed lines are the 700 hPa geopotential height contours (m). Red triangles indicate the locations of fire hot spots on July 5 from GOES-8 imagery. The black diamond shows the location of GSFC.

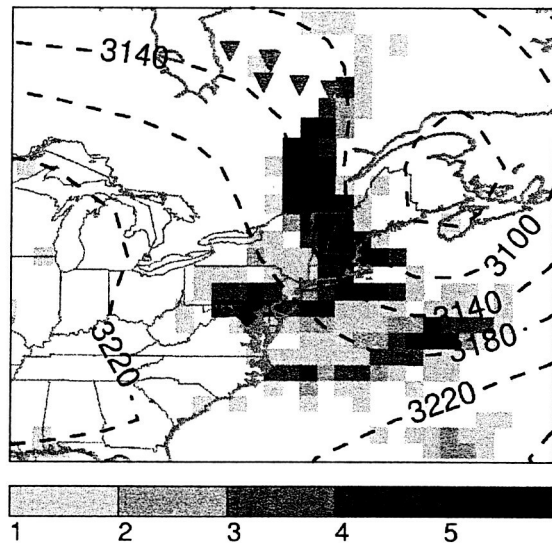


Figure 2. Vertical profile of aerosol at GSFC from (a) MPLNET and (b) the aerosol model. The white line is the MATCH PBLH. Note that the MPL did not collect data after about 1200 UTC on July 8 until late on July 9.

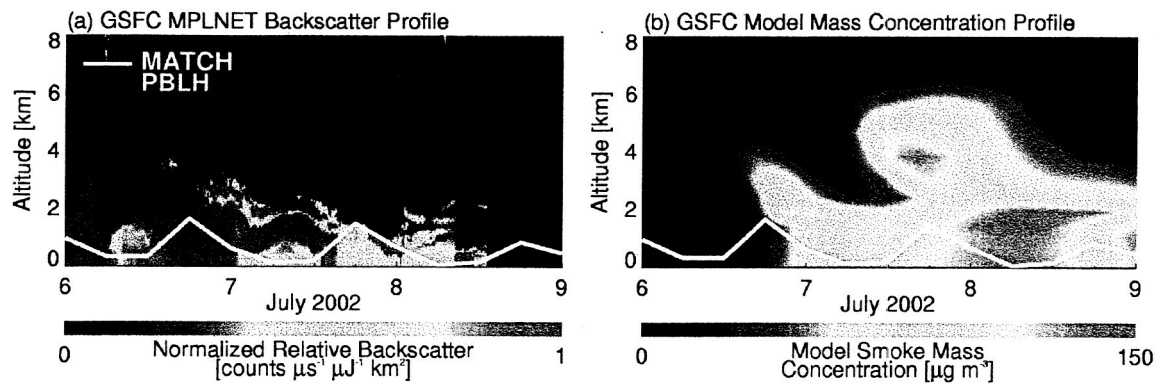


Figure 3. Two-day back trajectories from GSFC at the altitude of the observed peak MPLNET backscatter. The color shading of the line indicates the start time of the trajectory. The grey bars show the times the fires were active in Quebec as inferred from GOES-8 imagery. The black diamonds on each trajectory show the time at which the trajectory passed south of 51° N.

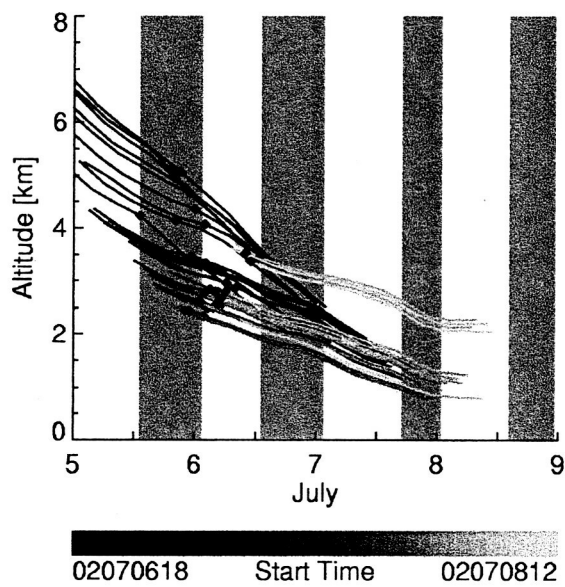


Figure 4. (a) EPA peak 1-hour average surface ozone mixing ratios on July 8, 2002.

Also shown are the modeled surface mass concentrations of smoke particles for the high-injection model runs (b) with and (c) without PBL entrainment.

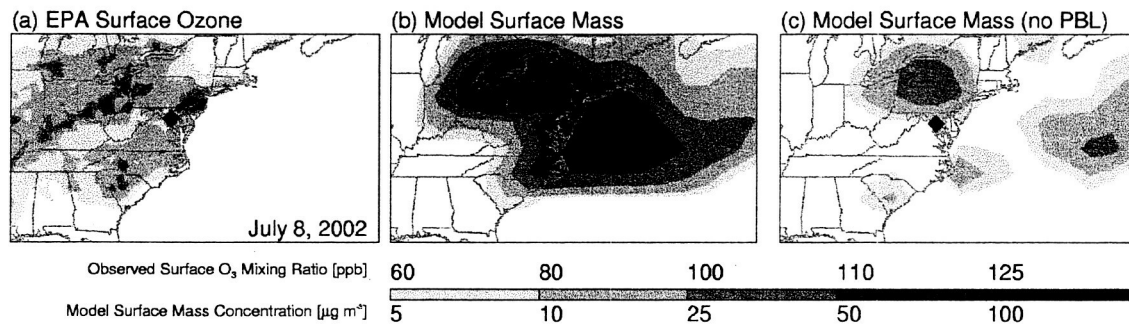


Figure 5. Observed and modeled aerosol (a) scattering and (b) absorption profiles at Easton on July 8, 2002. Shown are the Aztec profiles (diamonds) and the model profiles at 20:00 UTC (thick lines). Also shown are the MPLNET measurements of scattering and extinction at GSFC at 11:30 UTC (thin line).

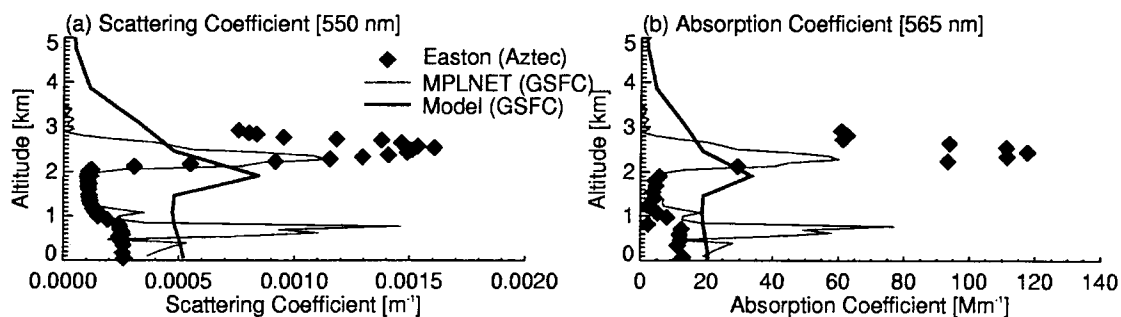


Table 1. Selected optical properties of the smoke aerosol observed near GSFC on July 8.

Shown are (a) the EP-TOMS retrieved AOT and single scatter albedo at 380 nm for the satellite footprint closest to GSFC, (b) the AERONET AOT at 380, 500, and 870 nm and single scatter albedo at 440 nm for the one hour period enclosing the EP-TOMS overpass, and (c) the Aztec AOT at 550 nm, single scatter albedo at 565 nm, and Ångström exponents for the Easton profile. Also shown are the model computed values both with and without coagulation.

	UTC	τ_{380}	τ_{500}	τ_{550}	τ	ω_0	$\tilde{A}_{450/550}$	$\tilde{A}_{450/700}$	$\tilde{A}_{550/700}$
EP-TOMS	16:11	2.77				0.89			
AERONET	15:42	2.09	1.75		0.75	$0.96 \pm$			
	16:27	2.32	1.96		0.84	0.03			
Aztec	20:00			$1.53 \pm$		$0.93 \pm$	$0.71 \pm$	$0.97 \pm$	1.18 ± 0.09
				0.34		0.01	0.10	0.09	
Model (coagulation)	20:00	2.55	2.17	2.00	1.00	0.96	0.80	1.06	1.28
Model (no coagulation)	20:00	2.34	1.64	1.44	0.62	0.96	1.41	1.58	1.73

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- 23 We acknowledge funding from XXX. BGD and LTM acknowledge support from Maryland Department of Environment and Mid-Atlantic and Northeast Visibility Union. Eric Kashiske and Brett Taubmann provided valuable comments to this manuscript.

Supplemental Online Material

Model Description

Our model (1,2) couples dynamical fields from the NCEP/NCAR reanalyses (3) and the NCAR Model for Atmospheric Transport and Chemistry (MATCH) (4) to the University of Colorado/NASA Ames Community Aerosol and Radiation Model for Atmospheres (CARMA) (5). CARMA is a bin-resolving microphysical model which solves the aerosol continuity equation for source, transport, removal, and transformational processes (e.g., coagulation). The reanalyses and MATCH provide wind, temperature, and pressure fields, as well as convective mass fluxes, precipitation rates, and eddy-diffusion coefficients for mixing in the PBL. These fields are archived globally every six hours at the resolution of the reanalyses (approximately $1.875^\circ \times 1.875^\circ$ in the horizontal and 21 vertical sigma levels from the surface to about 15 km). CARMA is run on a limited domain encompassing most of North America and the North Atlantic Ocean with the same grid spacing. The model is run from July 5 to 11 with an 1800 second time step and we linearly interpolate the input dynamical fields to the current time step.

The timing and locations of smoke emissions are determined from GOES-8 imagery of the fire hot spots (XXX). The locations of fire hot spots on July 5, 2002, are shown in Figure 1, and the active fire times are indicated by the grey bars in Figure 3. We assume that the fires are emitting at a constant rate while the fires are active. We assume a flux of $5 \times 10^{-8} \text{ kg m}^{-2} \text{ s}^{-1}$ per hot spot, which yields total emissions of about 1.5 Tg. The

injection altitude of the smoke particles is as described in the text.

We do not have a dataset to quantify the actual smoke emissions. A rough estimate was obtained by assuming typical land-surface characteristics and assuming 250,000 ha of land burned (6), where we assume average biofuel loadings of 17 Mg ha⁻¹ of above-ground carbon material (consumed with 26% efficiency) and 108 Mg ha⁻¹ of ground-level carbon material (consumed with 6% efficiency) (7). Assuming the carbon content is 45% of the total dry matter mass and applying a typical emission factor of 0.013 g PM_{2.5} per 1 g dry matter burned (8) we estimate total PM_{2.5} emissions from the fires to be about 0.10 Tg, which is about a factor of 10 smaller than the emissions used in the model. The emissions used in the model were tuned to give reasonable agreement between the model and observed aerosol optical properties at GSFC. Reducing the model emissions to 0.10 Tg results in a dramatic decrease in the AOT at GSFC so that the model greatly underestimates the observed AOT. We point out that there is considerable uncertainty in this emission estimate because we are applying average land spatial characteristics and assuming a smoke emission factor which may not be representative of the actual fires. Alternatively, because of numerical diffusion and the model's coarse spatial resolution we may require unrealistically large total emissions in order to get 200 km x 200 km grid cells to represent aerosol distributions which might be present on much smaller scales. Higher spatial resolution simulations will answer this question, but for now we are satisfied to get reasonably AOT agreement between the model and the observations.

The injected smoke particles are distributed across 16 size bins spaced logarithmically in radius between 0.01 – 1 μm . There are observations of coarse mode

particles with radius $> 1 \mu\text{m}$ in boreal forest fire plumes (9). These particles are usually ash, partially burned foliage, and lofted soil particles that make up a small fraction of the total particulate mass, have large fall speeds, and are less optically efficient than the fine mode particles. We neglect these large particles in this study of long-range transport of smoke, although they may be important to the properties of the smoke plume near the sources. We assume the smoke particles have a lognormal size distribution with a volume mode radius of $0.145 \mu\text{m}$ and a standard deviation of 2.00 (10). This initial particle size distribution is similar to a recent climatology of South American and African biomass burning aerosols determined from AERONET observations near major source regions (11). The density of the particles is assumed to be 1.35 g cm^{-3} (12).

Sedimentation, dry deposition, and precipitation scavenging rates are determined for each size bin (2). For some of our simulations we allow coagulation to modify the particle size distribution, which increases the mean radius of the particle size distribution at a rate approximately proportional to the air temperature, the square of the particle number concentration, and the inverse of the particle radius (13). Our algorithm (5) preserves the total aerosol volume but decreases the number concentration of particles as small particles stick together and grow larger. We assume a refractive index for the smoke particles of $m = 1.60 - 0.008i$, which is determined from AERONET retrievals of the smoke aerosol optical properties over GSFC on July 8, 2002 (14,15).

Uncertainties in the modeled aerosol optical properties result from assumptions made about the aerosol microphysics. We have not considered hygroscopic growth of the smoke aerosols during transport, although since the Aztec profiles show that the relative

humidity at the 2 – 3 km altitude of the smoke plume was less than 30% this process is probably unimportant to modifying the smoke optical properties. We also did not consider that the smoke in the plume may have mixed with clouds or other aerosol species, although because the plumes were mainly transported in an elevated layer they were probably largely unaffected by clouds or aerosols until after they had been intercepted by the PBL. Finally, our computed optical properties are sensitive to the choice of refractive index we made, which reflects the assumed composition of the aerosol.

Figure S1. SeaWiFS image of Canadian forest fires plumes over the northeastern U.S. on July 7, 2002. Image courtesy of...

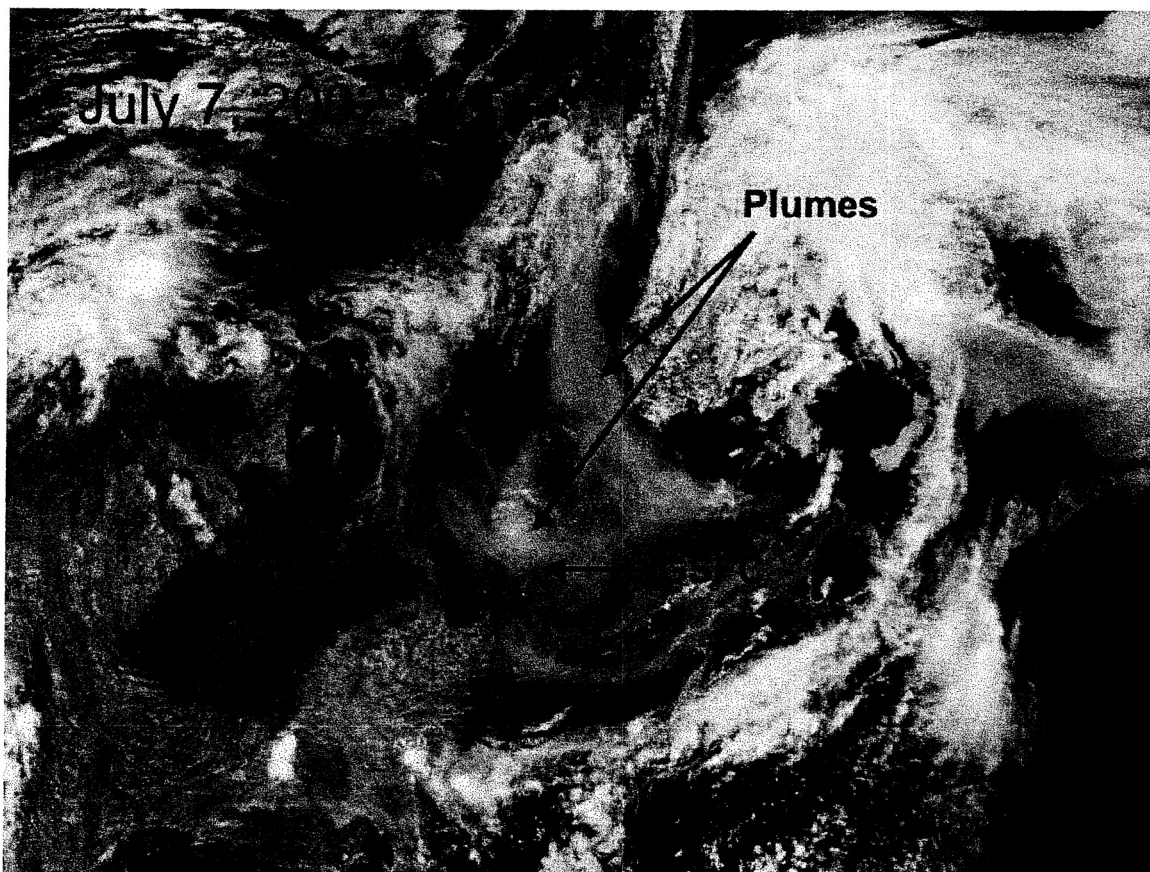


Figure S2. Two-day back-trajectories from GSFC. The shading of the trajectories indicates the initialization time of the trajectory, with initial times and altitudes consistent with Figure 3. The black diamond shows the location of GSFC. Red triangles are the GOES-8 fire locations on July 5, 2002. The dashed line shows the 51° N latitude, approximately the southern edge of the fire region.

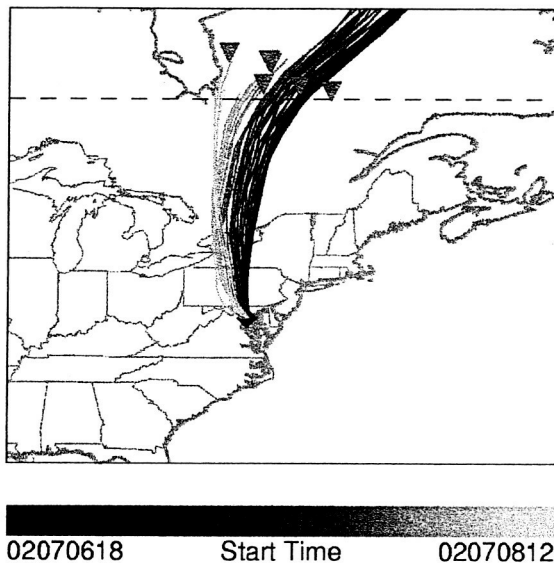


Figure S3. EPA peak 1-hour average ozone mixing ratios for July 6 – 9, 2002.

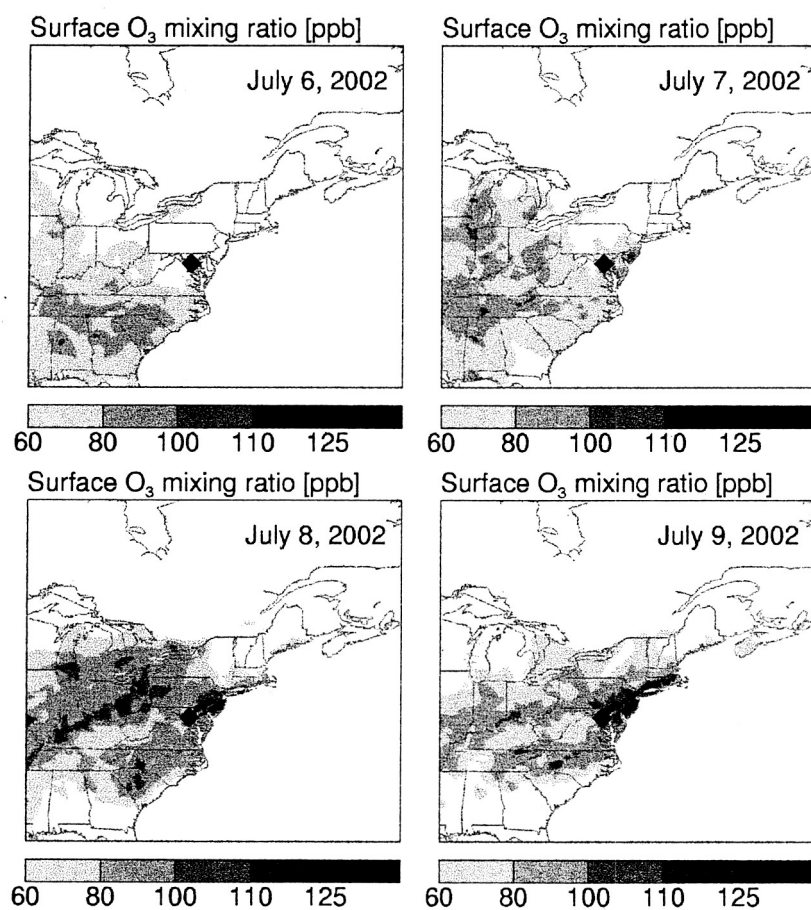


Figure S4. Observed and modeled aerosol vertical profiles. As in Figure 2 except that the model shown in (b) is from a run with PBL entrainment turned off.

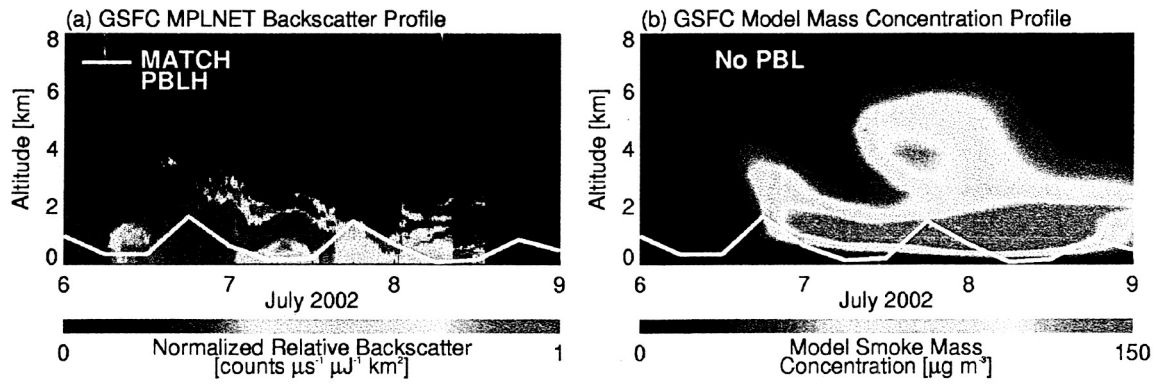
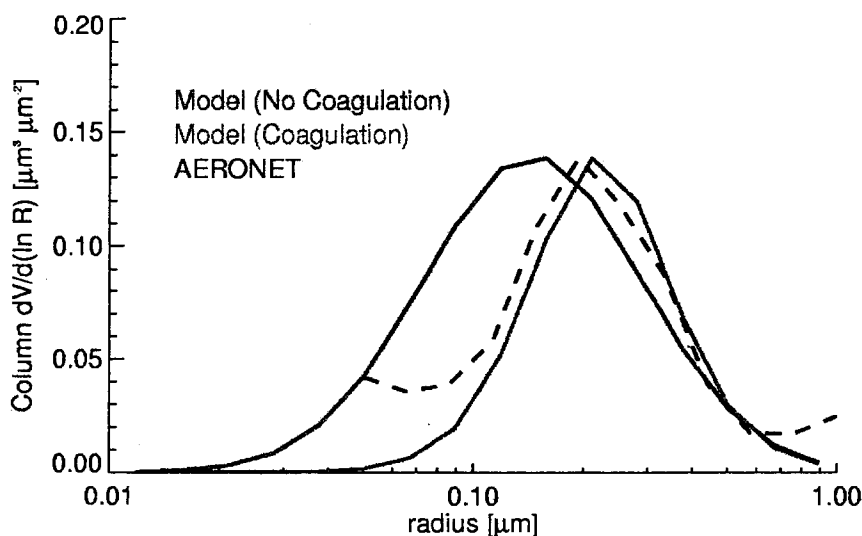


Figure S5. Observed and modeled aerosol particle size distribution at GSFC on July 8, 2002. The size distributions are integrated over the atmospheric column and normalized to the same peak value in the volume. Recall that the model does not contain anthropogenic aerosols in the PBL, which are seen by AERONET if present. Red line is the size distribution from the baseline model run with coagulation turned off. Blue line is from the baseline model run with coagulation turned on. Black dashed line is the AERONET retrieval (15).



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